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To cite this article: Yi Wang et al 2021 Environ. Res. Lett. 16 084061

View the article online for updates and enhancements.

### ENVIRONMENTAL RESEARCH LETTERS

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**OPEN ACCESS** 

RECEIVED 16 February 2021 REVISED

19 July 2021

ACCEPTED FOR PUBLICATION 21 July 2021

PUBLISHED 6 August 2021

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# Improved modelling of soil $NO_x$ emissions in a high temperature agricultural region: role of background emissions on $NO_2$ trend over the US

### Yi Wang<sup>1,2,\*</sup>, Cui Ge<sup>1,2,3</sup>, Lorena Castro Garcia<sup>1,2</sup>, G Darrel Jenerette<sup>4</sup>, Patty Y Oikawa<sup>5</sup> and Jun Wang<sup>1,2,\*</sup>

- Center for Global and Regional Environmental Research, Iowa Technology Institute, The University of Iowa, Iowa City, IA 52242, United States of America
- <sup>2</sup> Department of Chemical and Biochemical Engineering, The University of Iowa, Iowa City, IA 52242, United States of America
- <sup>3</sup> South Coast Air Quality Management District, Diamond Bar, CA 91765, United States of America
- <sup>4</sup> Department of Botany and Plant Sciences, University of California, Riverside, CA 92521, United States of America <sup>5</sup> Department of Farth and Environmental Sciences, California State University, Fact Pay, Henryard, CA 94542, United <sup>5</sup> Department of Farth and Environmental Sciences, California State University, Fact Pay, Henryard, CA 94542, United <sup>5</sup> Department of Farth and Environmental Sciences, California, State University, Fact Pay, Henryard, CA 94542, United <sup>5</sup> Department of Farth and Environmental Sciences, California, State University, Fact Pay, Henryard, CA 94542, United States, California, State University, Fact Pay, Henryard, CA 94542, University, Fact Pay, Henryard, CA 94542
- <sup>5</sup> Department of Earth and Environmental Sciences, California State University, East Bay, Hayward, CA 94542, United States of America
  <sup>\*</sup> Authors to whom any correspondence should be addressed.

E-mail: jun-wang-1@uiowa.edu and yi-wang-4@uiowa.edu

**Keywords:** soil NO<sub>x</sub> emissions, OMI NO<sub>2</sub>, high temperature, trend

Supplementary material for this article is available online

### Abstract

LETTER

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EPA reports a steady decline of US anthropogenic NO<sub>x</sub> emissions in 2005–2019 summers, while NO<sub>2</sub> vertical column densities (VCDs) from the OMI satellite over large spatial domains have flattened since 2009. To better understand the contributing factors to a flattening of the OMI NO<sub>2</sub> trends, we investigate the role of soil and lightning NO<sub>x</sub> emissions on this apparent disagreement. We improve soil NO<sub>x</sub> emissions estimates using a new observation-based temperature response, which increases the linear correlation coefficient between GEOS-Chem simulated and OMI NO<sub>2</sub> VCDs by 0.05–0.2 over the Central US. Multivariate trend analysis reveals that soil and lightning NO<sub>x</sub> combined emissions trends change from -3.95% a<sup>-1</sup> during 2005–2009 to 0.60% a<sup>-1</sup> from 2009 to 2019, thereby rendering the abrupt slowdown of total NO<sub>x</sub> emissions reduction. Non-linear inter-annual variations explain 6.6% of the variance of total NO<sub>x</sub> emissions. As background emissions become relatively larger with uncertain inter-annual variations, the NO<sub>2</sub> VCDs alone at the national scale, especially in the regions with vast rural areas, will be insufficient to discern the trend of anthropogenic emissions.

### 1. Introduction

 $NO_x$  (NO + NO<sub>2</sub>) emitted mainly from anthropogenic sources, wild fires, lightning, and soil plays a significant role in catalysing tropospheric ozone production and destruction and forming nitrate aerosols (Schindlbacher *et al* 2004, Seinfeld and Pandis 2016). National Emissions Inventory (NEI) reports a steady decrease of anthropogenic NO<sub>x</sub> emissions during 2005–2019 summers at a mean rate of -4.12% a<sup>-1</sup> over the contiguous US. Ozone Monitoring Instrument (OMI) tropospheric vertical column densities (VCDs), however, show a consistent decline until 2009, and a flattened trend afterwards up to 2015 (Jiang *et al* 2018). Silvern *et al* (2019) shows that the inconsistence between NEI and VCD trends during 2009–2017 reflect the variation of background emissions (soil and lightning sources). To date, the impact of soil  $NO_x$  emissions on the observational  $NO_2$  trends remains quantitively elusive.

Soil NO<sub>x</sub> emissions are an important source during summer (Vinken *et al* 2014, Huber *et al* 2020), and their strength is mainly influenced by fertilizer application, soil moisture, and soil temperature. Unusually high soil NO<sub>x</sub> emissions are observed over hightemperature agricultural regions in California where fertilized soil is a major source of NO<sub>x</sub> pollution (Oikawa *et al* 2015, Almaraz *et al* 2018). Moreover, nearly half of the enhancement of O<sub>3</sub> production that is related to rising temperature over the southeast US can be caused by the rising soil NO<sub>x</sub> emissions due to the increase of soil temperature (Romer *et al* 2018).

Here, we use satellite observations and a chemistry transport model (CTM) to improve estimates of soil  $NO_x$  emissions, and subsequently, to quantify the impact of soil  $NO_x$  on the slowdown of observed NO2 VCD reduction and examine the sources for inter-annual variation of total NO<sub>x</sub> emissions beyond the trend analysis. In various CTMs, soil  $NO_x$  emission estimates are often parameterized as a function of biome types, soil moisture and temperature, N-pulsing, N-deposition, and fertilizer maps (Yienger and Levy 1995, Steinkamp and Lawrence 2011, Hudman et al 2012, Rasool et al 2016, 2019). Two of uncertainties are the dependence of soil NO<sub>x</sub> emissions on soil temperatures especially in high temperature conditions and the simplification for soil temperature estimations. Notably, the emissions response to soil temperature in these CTMs is assumed to be flat after soil temperature is above 30 °C. Berkeley-Dalhousie Soil NO<sub>x</sub> Parameterization (BDSNP) is the state-of-the-art scheme, which is implemented in GEOS-Chem, a widely used CTM; we will improve it using a new observationbased relationship between soil surface NO<sub>x</sub> flux and soil temperature from Oikawa et al (2015) over the US in high temperature (above 30 °C) conditions. Furthermore, in existing GEOS-Chem, soil temperature as a key input variable to BDSNP is derived from 2 m air temperature through linear and empirical relationships that vary only with soil and canopy types (Yienger and Levy 1995). Hence, the soil temperature in the Modern-Era Retrospective analysis for Research and Applications, Version 2 (MERRA-2) dataset is used here directly to further refine BDSNP, thereby collectively formulating a new scheme, hereafter BDISNP, in which 'I' stands for Iowa.

Finally, the inconsistency between the trends of satellite-observed  $NO_2$  and anthropogenic  $NO_x$  emissions is investigated. With the improved soil emissions scheme, we quantitively attribute the slowdown of satellite-observed  $NO_2$  decreasing trend to different sectors and suggest that it is partly due to the co-variations of natural emissions and the trends thereafter among different sectors (such as soil and lightning).

### 2. GEOS-Chem modelling

# 2.1. BDISNP scheme with new function for high temperature

GEOS-Chem is a 3D chemistry transport model that is driven by meteorological re-analysis fields (GEOS-FP or MERRA-2). In its BDSNP scheme (Hudman *et al* 2012, Vinken *et al* 2014), the soil temperature response is identical to the exponential function ( $f(T) = e^{0.103T}$ ) for wet soil in YL95 (Yienger and Levy 1995) for temperature between 0 °C and 30 °C, and is constant once soil temperature is above 30 °C (blue line, figure 1). This study replace the old temperature function with a

new observation-based temperature function (orange line, figure 1) from Oikawa et al (2015), and the rest parameterizations are unchanged. In the range of 20 °C-40 °C, it is a cubic function of soil temperature  $(f(T) = -0.009T^3 + 0.837T^2 - 0.009T^3)$ 22.527T + 196.149). The value of the cubic function is close to the YL95 function in the range of 20 °C-23 °C, and it becomes larger than the exponential function as soil temperature increases from 23 °C to 40 °C. The cubic function is a factor of 1.4 and 2.7 larger than the exponential function at 30 °C and 40 °C, respectively. When soil temperature is above 40 °C, temperature function is set as the value of the cubic function at 40 °C. The YL95 exponential function is now used for soil temperature between 0 °C and 20 °C.

### 2.2. Experimental design

We use version 12.7.2 GEOS-Chem driven by MERRA-2 meteorological fields to simulate soil NO<sub>x</sub> emissions and atmospheric NO2 concentrations over the US, with a spatial resolution of  $0.5^{\circ} \times 0.625^{\circ}$ and 47 vertical layers from surface to stratopause in summer (June, July and August) during 2005–2019. The boundary conditions required by the nested simulations are provided by global  $2^{\circ} \times 2.5^{\circ}$  simulations (Wang et al 2004). Anthropogenic emissions are distributed using the NEI2011 inventory scaling to different years according to national annual totals (EPA 2021). Open fire emissions are from the fourthgeneration Global Fire Emission Database (GFED4) daily emission inventory (Giglio et al 2013). Lightning NO<sub>x</sub> emissions are calculated using the lightning flash densities and convective-cloud depths that are derived from the MERRA-2 data and distributed following the monthly redistribution factors generated using the Lightning Imaging Sensor and Optical Transient Detector (LIS/OTS) high resolution monthly climatology (Murray et al 2012).

Three sensitivity experiments with different treatments of soil  $NO_x$  emissions are used to explore the impacts of the updated scheme (BDISNP) on soil  $NO_x$  emissions and atmospheric  $NO_2$  concentrations. The Control and  $T_{new}$  runs use BDSNP and BDISNP, respectively, while soil  $NO_x$  emissions are turned off in the NO\_SOIL experiment. Following Silvern *et al* (2019), the soil  $NO_x$  emissions are reduced by 50% for the control and  $T_{new}$  runs, based on the previous comparison between GEOS-Chem simulations and OMI  $NO_2$  VCDs showing the overestimation of soil  $NO_x$  emissions using BDSNP (Vinken *et al* 2014).

# 3. OMI and surface NO<sub>2</sub> and GEOS-Chem data processing

Tropospheric NO<sub>2</sub> VCDs from the version 4.0 NASA OMI NO<sub>2</sub> L2 standard product (OMINO2) (Lamsal *et al* 2021) at a spatial resolution of  $13 \times 24 \text{ km}^2$ 





at nadir are used to evaluate the GEOS-Chem simulations. The product has a swath of 2600 km and an equator-crossing time at 13:45, hence providing contiguous global coverage daily (Levelt *et al* 2006). Retrievals are error prone to large Solar Zenith Angles (SZA) and large View Zenith Angles (VZA), high cloud contamination, and snow cover; thus only data with SZA < 75°, VZA < 65°, effective cloud fraction < 0.3, terrain reflectivity < 0.3 are used (Wang *et al* 2020a). Moreover, the pixels that are affected by row anomaly (Schenkeveld *et al* 2017) or not flagged as high quality are removed.

In situ surface NO<sub>2</sub> measurements during 2005–2019 are available at EPA Air Quality System (AQS), and most of these observational sites are located in urban regions. Despite of systematic positive bias due to the measuring method (Steinbacher *et al* 2007), the measurement are still useful for model evaluations (Wang *et al* 2020b).

For proper comparison, GEOS-Chem NO<sub>2</sub> simulations are sampled according to OMI observing time and converted to the tropospheric VCDs using the scattering weights, tropospheric air mass factors (AMFs), and tropopause pressure from the OMINO2 product (Wang *et al* 2020a). Comparison of GEOS-Chem surface NO<sub>2</sub> with surface *in situ* data from EPA is spatially and temporally paired, following the method in Wang *et al* (2020b).

### 4. Approaches for trend analysis

A statistical framework with treatment of covariance of emissions in each sector is conceived for the trend analysis and attribution, and to quantitively link the linear trend and inter-annual non-linear variation of  $NO_x$  emissions from these sectors to the variance of total NO<sub>x</sub> emissions, hence explaining the atmospheric NO<sub>2</sub> trend observed by satellite. Firstly, total NO<sub>x</sub> emissions consist of the trend parts and interannual variability parts of each sectors shown as:

$$E_{i}^{t} = \sum_{j=1}^{N} E_{i,j}^{tr} + \sum_{j=1}^{N} E_{i,j}^{vr} = \sum_{j=1}^{N} c_{j} + \sum_{j=1}^{N} b_{j} \cdot i + \sum_{j=1}^{N} E_{i,j}^{vr}$$
(1)

where *E* represents  $NO_x$  emissions, t, tr, and vr are short for total, trend, inter-annual variability, respectively, *i* and *j* are indexes of year and sector, respectively, *N* is total number of sectors including anthropogenic, soil, lightning, and fires, *c* and *b* are intercept and slope (trend) of *E*<sup>tr</sup>, respectively, which are derived through linear regression.

Following past studies, the relative trends of total  $(b^{\text{rel},t})$  and sectoral  $(b_j^{\text{rel}})$  NO<sub>x</sub> emissions since 2005 are the focus of this study; their relationship is shown as:

$$b^{t} = \sum_{j=1}^{N} b_{j}$$
$$(b^{t} \div E_{2005}^{t}) \cdot E_{2005}^{t} = \sum_{j=1}^{N} (b_{j} \div E_{2005,j}) \cdot E_{2005,j}$$
$$b^{\text{rel},t} \cdot E_{2005}^{t} = \sum_{j=1}^{N} b_{j}^{\text{rel}} \cdot E_{2005,j}$$
$$b^{\text{rel},t} = \sum_{j=1}^{N} b_{j}^{\text{rel}} \cdot \frac{E_{2005,j}}{E_{2005}^{t}}$$
(2)

Furthermore, we show that the variance of total  $NO_x$  emission can be divided into: (a) the covariances between sectoral trend parts and total  $NO_x$  emissions and, (b) the covariances between

sectoral inter-annual variability parts and total NO<sub>x</sub> emissions:

$$\operatorname{Var}\left(\boldsymbol{E}^{\mathsf{t}}\right) = \operatorname{Var}\left(\sum_{j=1}^{N} \boldsymbol{E}_{j}\right) = \operatorname{Var}\left(\sum_{j=1}^{N} \boldsymbol{E}_{j}^{\mathsf{tr}} + \sum_{j=1}^{N} \boldsymbol{E}_{j}^{\mathsf{vr}}\right)$$
$$= \operatorname{Cov}\left(\boldsymbol{E}^{\mathsf{t}}, \sum_{j=1}^{N} \boldsymbol{E}_{j}^{\mathsf{tr}}\right) + \operatorname{Cov}\left(\boldsymbol{E}^{\mathsf{t}}, \sum_{j=1}^{N} \boldsymbol{E}_{j}^{\mathsf{vr}}\right)$$
$$= \sum_{j=1}^{N} \operatorname{Cov}\left(\boldsymbol{E}_{j}^{\mathsf{tr}}, \boldsymbol{E}^{\mathsf{t}}\right) + \sum_{j=1}^{N} \operatorname{Cov}\left(\boldsymbol{E}_{j}^{\mathsf{vr}}, \boldsymbol{E}^{\mathsf{t}}\right)$$
(3)

where Var and Cov represent variance and covariance operator, respectively, E is a data vector consisting time series of the NO<sub>x</sub> emissions data. Hence, the contribution of sectoral linear trend to the variance of total NO<sub>x</sub> emissions can be computed as  $Cov(E_j^{tr}, E^t)/Var(E^t)$ , while  $Cov(E_j^{vr}, E^t)/Var(E^t)$ is the part of total emission variance explained by the non-linear sectoral inter-annual variability. Subsequently, the trend part of the variance of total NO<sub>x</sub> emissions, Var( $E^{tr,t}$ ), is essentially the sum of covariances between sectoral emission trend ( $E_j^{tr}$ ) and total emission trend ( $E^{tr,t}$ ):

$$\operatorname{Var}\left(\boldsymbol{E}^{\mathrm{tr},\mathrm{t}}\right) = \operatorname{Var}\left(\sum_{j=1}^{N} \boldsymbol{E}_{j}^{\mathrm{tr}}\right) = \operatorname{Cov}\left(\boldsymbol{E}^{\mathrm{tr},\mathrm{t}},\sum_{j=1}^{N} \boldsymbol{E}_{j}^{\mathrm{tr}}\right)$$
$$= \sum_{j=1}^{N} \operatorname{Cov}\left(\boldsymbol{E}_{j}^{\mathrm{tr}},\boldsymbol{E}^{\mathrm{tr},\mathrm{t}}\right)$$
(4)

Hence, for each sector *j*, its contribution to the linear trend of total emissions can be computed as  $\operatorname{Cov}\left(\boldsymbol{E}_{j}^{\mathrm{tr}}, \boldsymbol{E}^{\mathrm{tr}, \mathrm{t}}\right) / \operatorname{Var}\left(\boldsymbol{E}^{\mathrm{tr}, \mathrm{t}}\right)$ .

### 5. Results

#### 5.1. Evaluation of BDISNP

 $NO_x$ T<sub>new</sub> produces soil emissions of 201 Gg N summer<sup>-1</sup> over the US, which is 18.2% higher than that from Control. Moreover, during the hot days when daily mean soil temperatures are greater than 30 °C, T<sub>new</sub> has 32.6% more soil NO<sub>x</sub> emissions than Control. Large soil NO<sub>x</sub> emissions from Control (figure 2(a)) are mainly over Central US and California, where agricultural activity is intense. Compared with Control, Tnew simulates the increase of soil  $NO_x$  emissions up to 120% mainly in south of  $40^{\circ}$ N (figure 2(c)) where soil temperature is high and consequently, the new temperature response function is important. In north of 40°N, up to 47% decrease of soil NO<sub>x</sub> emissions exists (figure 2(c)), which is caused by that MERRA-2 soil temperature (T<sub>new</sub>) is lower than that calculated from 2 m temperature using linear and empirical approach (Control) in these regions.

Evaluation of GEOS-Chem simulations with OMI tropospheric VCDs mainly focuses on temporal linear correlation coefficients between them, which are highly relevant to the linear trend analysis. Compared with the Control run (figure 3(a)),  $T_{new}$  shows the increase of linear correlation coefficient (R) in the range of ~0.05 to ~0.2 over the Central US (black box in figure 3(b)), where fertilizer activities are intense and soil NO<sub>x</sub> emissions play a significant role. The small decrease (usually less than 0.04) of R is also found over the North Dakota, northern Minnesota, and central Texas. Overall, BDISNP improves on BDSNP in terms of R values over the US when evaluating with OMI tropospheric NO<sub>2</sub> VCDs.

The improvement of R over the Central US is caused by the better simulations of tropospheric NO<sub>2</sub> VCDs over high soil temperature conditions (>30 °C). Figure 3(c) show the changes of tropospheric NO2 VCDs from OMI and the GEOS-Chem simulations as soil temperature increases over the Central US. When soil temperature is less than 30 °C, the difference between Control and T<sub>new</sub> is negligible (results not shown). In the soil temperature range of 30 °C to 40 °C, the tropospheric NO<sub>2</sub> VCD from  $T_{new}$ is larger than Control, and the difference increases as soil temperature rises (figure 3(c)). T<sub>new</sub> has larger positive bias than the Control (figure 3(c)), but the positive bias is not caused by the parameterization for emissions dependence on temperature; even if soil NO<sub>x</sub> emissions are excluded in the GEOS-Chem simulation (NO\_SOIL), modeled NO<sub>2</sub> VCD is still larger than OMI observation (figure 3(c)). The positive bias likely originates from the overestimation of other terms in  $NO_x$  emission schemes (that this study does not revise) or other error sources in GEOS-Chem. Furthermore, the uncertainties from NO<sub>2</sub> slant column fitting, tropospheric-stratospheric separation (with year to year variability and is most relevant for rural studies), and air mass factor calculation (Bucsela et al 2013) all contribute to the uncertainty of OMI tropospheric NO<sub>2</sub> VCD, which is estimated as 35% (Lamsal et al 2021), and all the bias in T<sub>new</sub> is within the uncertainty envelope (shaded area in figure 3(c)) of OMI observational error (Lamsal et al 2021). We assume that NO<sub>2</sub> VCD difference between OMI and the GEOS-Chem simulations are independent of soil temperature, correcting the GEOS-Chem systematic bias by subtracting the difference for all conditions with soil temperature above 30 °C. With this non-temperature dependent bias correction (figure 3(c)), T<sub>new</sub> well captures the increase of OMI tropospheric NO2 VCD observations as a function of soil temperature, while both Control and NO\_SOIL underestimate the characteristics (figure 3(d)); the Normalized Centered Root Mean Square Error (NCRMSE) (Wang et al 2020a) is 0.089 for  $T_{new}$ , which is much smaller than 0.266 and 0.421 for Control and NO\_SOIL, respectively (figure 3(d)). As to the decrease of



R(T<sub>new</sub>)-R(Control) R(Control) 50°N 40°N 30°N (b) (a) 120°W 100°W 80°W 120°W 100°W 80°W 0.2 0.4 0.6 0.8 1.0 0.0 -0.2 -0.1 0.0 0.1 0.2 1e15 1e15 3.0 2.1 Bias-corrected NO<sub>2</sub> VCD [molec cm<sup>-2</sup>] OMI 2.0 Control(0.266) 2.5  $T_{new}(0.089)$ NO<sub>2</sub> VCD [molec cm<sup>-2</sup>] 1.9 NO\_SOIL(0.421) 2.0 1.8 1.5 1.7 1.6 1.0 OMI Control(0.99) 1.5 0.5  $T_{new}(1.00)$ (d) (c) NO\_SOIL(0.97) 1.4 0.0 . 34 36 38 32 36 38 40 32 34 40 30 30 Soil temperature [°C] Soil temperature [°C]







R over North Dakota/Minnesota (orange box in figure 3(b)) and Texas (green box in figure 3(b)), figure S1 (available online at stacks.iop.org/ERL/16/ 084061/mmedia) shows that OMI NO<sub>2</sub> VCDs are almost constant in temperature ranges of 20 °C– 30 °C and 30 °C–38 °C for North Dakota/Minnesota and Texas, respectively, while all GOES-Chem simulations (Control,  $T_{new}$ , and NO\_SOIL) show NO<sub>2</sub> VCDs increase as soil temperature rises. As GEOS-Chem cannot properly simulate NO<sub>2</sub> VCDs in the two regions, the new scheme even renders a decrease of R.

GEOS-Chem surface NO<sub>2</sub> concentrations from Control and  $T_{new}$  are also evaluated with EPA *in situ* observations. Surface NO<sub>x</sub> concentrations between  $T_{new}$  and EPA have R values ~0.02–0.03 larger than of that from the Control almost each summer (figure 4(a)), while NCRMSE and normalized mean square error (NMSE) decrease in the range of 2.0%– 5.3% (figure 4(b)) and 11.1%–17.9% (figure 4(c)), respectively, each summer.

# 5.2. Quantifying background contributions to $\mathrm{NO}_{\mathrm{x}}$ variations

Figures 5(a) and (b) show the relative trends  $(b^{rel})$ since 2005 of NO<sub>x</sub> emissions from T<sub>new</sub> and tropospheric NO2 VCDs from OMI and the GEOS-Chem simulations in contagious US, respectively. Total NO<sub>x</sub> emissions show a steady decline during 2005–2019, primarily reflecting the trend of its dominant source—the anthropogenic NO<sub>x</sub> emissions (figure 5(a)). Starting from 2009, however, the decline trend of total NO<sub>x</sub> emissions becomes less than half of that during 2005-2009 (figure 5(a)). OMI tropospheric NO<sub>2</sub> VCDs observe a steady decline from 2005 to 2009 and a flattening afterward, supporting the existence of the turning point of total  $NO_x$  trend in 2009 as simulated by  $T_{new}$  (figure 5(b)), and the stronger slowdown of NO2 VCDs reduction in Central US than in Western and Eastern US implies that NO<sub>x</sub> emissions from natural sources contribute to the slowdown (figure S2). Therefore, the fidelity of T<sub>new</sub> in simulating the NO<sub>x</sub> VCD variation (section 5.1 and figure 5(b) enables further quantitative attribution of relative trend of total NO<sub>x</sub> emissions to role of emissions from each sector.

The relative trend since 2005 of total NO<sub>x</sub> emissions is -2.75% a<sup>-1</sup> (figure 5(a)); it equals the sum of the relative trends of every sector weighted by the ratio of sectoral to total NO<sub>x</sub> emissions in 2005 (equation (2)). NO<sub>x</sub> emissions from fires sector have the largest positive relative trend of 6.49%  $a^{-1}$  (figure 5(a)), but its contribution to the relative trend of total  $NO_x$  emissions is negligible, as the percentage of fires sector to total  $NO_x$  emissions is only 0.87% in 2005 and 1.89% during 2005-2019. The contributions of soil, lightning, and anthropogenic sector to total NO<sub>x</sub> emissions in 2005 are 11.15%, 16.10%, and 71.88%, respectively. Hence the relative trend of total NO<sub>x</sub> emissions is mainly affected by these three sources. The relative trend of anthropogenic NO<sub>x</sub> emissions is -4.12% a<sup>-1</sup>, while soil sector has a weaker downward trend of -1.44%a<sup>-1</sup> and lightning sector has an upward trend of 1.94%  $a^{-1}$  (figure 5(a)). As a result, despite the large steady decrease of NO<sub>x</sub> emissions from anthropogenic source, total NO<sub>x</sub> emissions trend is weakened by the impacts of the increase of lightning emissions and the weaker decrease of soil emissions to reach -2.75% a<sup>-1</sup> overall during 2005–2019.

The trends relative to 2005 in the periods of 2005–2009 and 2009–2019 for different  $NO_x$  emissions sources are calculated (table 1) to investigate the abrupt slowdown of total  $NO_x$  emissions reduction starting in 2009. The relative decreasing trend in total NO<sub>x</sub> emissions is -5.19% from 2005 to 2009, which is 120% stronger than that during 2009–2019. The relative trend in anthropogenic NO<sub>x</sub> emissions is only 53% stronger during 2005-2009 than that from 2009 to 2019. The much larger slowdown of the total NO<sub>x</sub> emissions reduction than that of anthropogenic source is caused by the difference between the two periods for relative trends in soil and lightning NO<sub>x</sub> emissions. The relative trends in lightning and soil sources during 2005–2009 are -1.15% a<sup>-1</sup> and -8.00% a<sup>-1</sup>, respectively, and their weighted mean is -3.95% a<sup>-1</sup>; by contrast, their combination shows an upward relative trend of 0.60%  $a^{-1}$  from 2009 to 2019. The change of the directions of the relative trends largely contribute to the slowdown of total NO<sub>x</sub> emissions reduction. The temporal variation of



**Figure 5.** Relative trends since 2005 of (a) NO<sub>x</sub> emissions by sectors from  $T_{new}$ , (b) tropospheric NO<sub>2</sub> VCDs derived from OMI and the GEOS-Chem Control,  $T_{new}$ , and NO\_SOIL simulations in contiguous US. (c) Is the decomposition of the variance of total NO<sub>x</sub> emissions (black dash line) according to equation (3) with each bar showing the covariance of component and total NO<sub>x</sub> emissions. (d) Is the decomposition of the variance of the trend in total NO<sub>x</sub> emissions (black dash line) according to equation (4) with each bar showing the covariance of component trend and trend in total NO<sub>x</sub> emissions.

Table 1. The NO<sub>x</sub> emissions trends relative to the start year of various time periods for difference sources (units:  $\% a^{-1}$ ).

	2005–2019	2005–2009	2009–2019
Total	$-2.75\pm0.20$	$-5.19\pm0.84$	$-2.34\pm0.32$
Anthropogenic	$-4.12\pm0.12$	$-5.66\pm0.05$	$-3.71\pm0.11$
Lightning	$1.94\pm0.82$	$-1.15 \pm 3.44$	$2.48 \pm 1.47$
Soil	$-1.44\pm1.21$	$-8.00\pm4.96$	$-1.82\pm2.11$
Fires	$6.49\pm 6.01$	$-5.01\pm21.82$	$14.74\pm10.66$

soil  $NO_x$  emissions are mainly affected by soil temperature and soil moisture rather than the change of fertilizer (S1 in supplement).

In addition to linear trends, we investigate the contribution of sectoral emissions to the inter-annual non-linear variations of NO<sub>x</sub> emissions following equation (3). The covariance between anthropogenic trend part and total NO<sub>x</sub> emissions is 55 123 (Gg N)<sup>2</sup> (gigagram N squared), which is slightly larger than the variance of total NO<sub>x</sub> emissions of 54 877 (Gg N)<sup>2</sup> (figure 5(c)). The sum of covariances

between sectoral trends and total NO<sub>x</sub> emissions is 51 229 (Gg N)<sup>2</sup>, suggesting that the linear trend among each sector can explain 93.4% of the variance of total NO<sub>x</sub> emissions, while the non-linear interannual variability explain 6.6% (figure 5(c)). Similarly, following equation (4), we find that the contribution of anthropogenic, soil, lightning, and fire sources to the trend of total NO<sub>x</sub> emissions in terms of variance are 107.60%, 5.84%, -11.35%, and 2.09%, respectively, in contagious US, and these values are 103.9%, 11.6%, -15.8%, 0.2% in Central US (defined in figure S2(a)). Thus, the non-linear inter-annual variability also plays an important role, and its contribution is expected to increase as anthropogenic  $NO_x$  emissions decrease over time.

### 6. Discussion and conclusions

We revised Berkeley-Dalhousie Soil NO<sub>x</sub> Parameterization (BDSNP) in GEOS-Chem by (a) allowing the increase of soil NO<sub>x</sub> emission with temperature up to 40 °C (instead of 30 °C in BDSNP) based on in situ observations, and (b) using the soil temperature directly from MERRA-2 (instead of indirectly deriving soil temperature from air temperate as in implemented in standard GEOS-Chem), hence forming the new BDSINP (where 'I' stands for Iowa) scheme. Compared to the GEOS-Chem simulation with BDSNP, using BDSINP overall yields better agreement with OMI NO<sub>2</sub> climatology; the linear correlation coefficients between simulated and OMIobserved tropospheric NO<sub>2</sub> VCDs during 2005–2019 summer increase by 0.05-0.2 over the Central US, where soil  $NO_x$  emissions dominate. Furthermore, GEOS-Chem with BDSINP renders a tropospheric NO<sub>2</sub> VCDs response function to temperature that is more consistent with the counterpart independently derived from a combined analysis from OMI and MERRA-2 data.

With its improved fidelity, GEOS-Chem simulation with BDSINP is used to disentangle the covariance among different NO<sub>x</sub> emission sectors toward understanding the slowdown of tropospheric NO2 VCD reduction during 2009-2019. A statistical framework is proposed to decompose the variance of total NO<sub>x</sub> emission into linear and non-linear parts, and further offer a quantitative approach to attribute each part to the linear and non-linear variation of emissions from each sector. GEOS-Chem simulation with BDSINP is able to capture the linear trend of OMI tropospheric NO2 VCDs, both showing a decreasing trend during 2005-2009 summer and a level-off after 2009. Indeed, in the total variance of total NO<sub>x</sub> emissions in 2005–2019,  $\sim$ 93.4% can be explained by linear trends from each sector, in which anthropogenic, soil, lightning, and fire sources contribute 107.60%, 5.84%, -11.35%, and 2.09%, respectively; the negative contribution of lightning emissions means they are negatively correlated with total  $NO_x$  emissions, hence dampening the trend in total  $NO_x$  emissions. The remaining 6.6% is due to non-linear inter-annual variations of each sector and their co-variations among each sector, and such nonlinear inter-annual variations is expected to increase in the next decade as anthropogenic emissions continue to decline while natural variations due to fires, lightning, and soil are expected to increase. Due to the impact of soil and lightning emissions on the total trend and the contribution of non-linear inter-annual variations to the variation of total NO<sub>x</sub> emissions, a

linear trend derived from the satellite observations in rural areas would be difficult to discern the trend of anthropogenic emissions.

This study has quantified that NO<sub>x</sub> natural sources (soil and lightning) contribute to the slowdown of NO<sub>2</sub> VCDs reduction, although the roles of anthropogenic source and atmospheric chemistry could be important and should be investigated in the future research. We have simulated the slowdown of NO<sub>2</sub> VCDs reduction starting at 2009 with the input of the EPA's emission inventory that presented the steady (and more or less linear) reduction of anthropogenic NO<sub>x</sub> emissions. In the past studies, Jiang et al (2018) reported another inventory that showed the slowdown of the reduction of anthropogenic NO<sub>x</sub> emissions due to 'the growing relative contribution of industrial, area, and off-road mobile sources of emissions, decreasing relative contribution of on-road gasoline vehicles, and slower than expected decreases in on-road diesel NO<sub>x</sub> emissions', and used it to explain the slowdown of the reduction of atmospheric NO<sub>2</sub>; in contrast, Silvern *et al* (2019) showed that both trends of two inventories from EPA and Jiang et al (2018) are within the uncertainty of AQS surface NO2 trends. Therefore, which anthropogenic inventory is more realistic is still open to discussion. However, this study showed quantitively the importance of soil and lightning  $NO_x$  in explaining the slowdown starting at 2009 (other than in the years prior to or after 2009), and pointed out that the linear trend only partially explains the variance of atmospheric NO<sub>2</sub> with time. While chemistry and anthropogenic emissions both play a role in the slowdown (as pointed by Laughner and Cohen (2019) and Jiang et al (2018)), this study reveals that the importance of soil emissions should not be overlooked toward a better understanding of the slowdown of atmospheric NO2 reduction, especially over the Central US where the slowdown is more significant than Eastern or Western U.S. (figure S2). An improved soil  $NO_x$ emission scheme for the Central U.S., as developed in this study, can be used to further detangle the relative role of chemistry, anthropogenic emissions, and natural source emissions in the change of atmospheric NO<sub>2</sub> with time.

### Data availability statement

The data that support the findings of this study are openly available at the following URL/DOI: https://disc.gsfc.nasa.gov/datasets/OMNO2\_003/summary?keywords=omi%20no2 (OMI NO2 product), https://aqs.epa.gov/aqsweb/airdata/down load\_files.html (EPA *in situ* NO2), http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem\_12#12.2.0 (GEOS-Chem source code and all required emission inventories (including GFED4 and NEI2011) other inputs), www.epa.gov/air-emissi ons-inventories/air-pollutant-emissions-trends-data

(EPA national annual anthropogenic NOx emissions totals).

### Acknowledgments

This work was supported by NASA Atmospheric Modeling and Analysis program (Grant Nos. NNX17AF77G and 80NSSC19K095), National Science Foundation (Grant No. DEB-1656062), USDA (Award No. 2019-67021-29227), and NOAA (Award No. NA19OAR4310178).

### ORCID iD

Jun Wang bhttps://orcid.org/0000-0002-7334-0490

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